

Two *p*-tert-butylthiacalix[4]arene-supported bismuth cluster compounds for ultrahigh efficiency detection of trace-level Cd²⁺ and Pb²⁺

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ABSTRACT

Heavy metal ions (HMIs) in the environment pose a significant threat to both environmental quality and human health. Consequently, monitoring HMIs concentrations is an essential yet challenging endeavor. Electrochemical sensing emerges as an effective method for detecting HMIs due to its high reliability and sensitivity. According to the hard/soft acid/base theory, porous metal-organic complexes containing S, O or N hybrid atoms provide the binding sites that facilitate the trapping of HMIs within cavities during the electrochemical deposition process, thereby enhancing detection sensitivity. In this study, we employed *p*-tert-butylthiacalix[4]arene ligands (H₄TC4A) to rationally design two Bi(III) clusters compounds: [Bi₄(TC4A)₂(μ₄-O)]·2NO₃·3DMF·3MeOH (**Bi₄-TC4A**) and [Bi₈(TC4A)₅(en)₂(μ₃-OH)₂(μ₄-OH)₂]·5DMF·2MeOH (**Bi₈-TC4A**). Furthermore, we constructed two Bi(III)-based electrochemical sensors, **Bi₄-TC4A/GCE** and **Bi₈-TC4A/GCE**, for the detection of Cd²⁺ and Pb²⁺ ions. Compared to **Bi₄-TC4A**, **Bi₈-TC4A** contains eight Bi(III) cations and five sulfur-containing TC4A⁴⁻ ligands, which significantly enhance the enrichment and adsorption of HMIs for electrochemical sensing applications. Notably, the linear ranges for Cd²⁺ and Pb²⁺ on the **Bi₈-TC4A/GCE** were found to be 0.01–1.1 mg·L⁻¹ and 0.01–1.4 mg·L⁻¹ with the detection limits of 1.4 and 1.7 μg·L⁻¹, respectively, which are lower than those specified by World Health Organization guidelines. This work offers valuable insights into utilizing well-designed Bi(III) cluster-based sensors for highly sensitive electrochemical detection of HMIs with low detection limits.

Environmental implication

Cd²⁺ and Pb²⁺ are highly toxic heavy metal pollutants that pose significant risks to the environment and human health. These contaminants can degrade soil and water quality, disrupt ecosystems, and accumulate in the food chain. In response to the urgent need for a rapid and accurate method for detecting HMIs in water and soil, this study has designed and synthesized two bismuth cluster compounds. These compounds effectively enrich HMIs by leveraging the cavity structure created by calixarene, enabling quantitative detection of Cd²⁺ and Pb²⁺.

1. Introduction

In recent years, environmental issues have garnered significant attention [1,2]. Among these concerns, heavy metal ions (HMIs) are particularly problematic as they are non-biodegradable and tend to accumulate in living organisms, potentially leading to severe health complications and physiological disorders [3–5]. Therefore, HMIs are recognized as a major source of global biosphere pollution [6,7]. In this context, cadmium ions (Cd²⁺), a highly toxic HMI, has fearful effects on kidney, liver, nerve system and cardiovascular health [8,9]. Lead ions (Pb²⁺) represent another critical concern due to their detrimental effects on human health, which can cause damage to the circulatory system, nervous system and immune response even at low concentrations [10,11]. Consequently, there is an urgent need for sensitive and selective

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detection methods for Cd^{2+} and Pb^{2+} that are both cost-effective and convenient [12–15]. Usually, traditionally employed analytical methods such as atomic absorption spectrometry (AAS) [16], atomic emission spectrometry (AES), and inductively coupled plasma-mass spectrometry (ICP-MS) have been utilized for the determination of HMIs [17,18], however, these methods often require expensive instrumentation and skilled personnel. In contrast, electrochemical square wave anodic stripping voltammetry (SWASV) offers advantages including straightforward operation, low cost, portability convenience, high sensitivity, and selectivity [19,20]. Thus, it becomes an alternative low-cost analytical approach for the electrochemical detection of Cd^{2+} and Pb^{2+} [21,22].

Bismuth (Bi)-based films serve as environmentally friendly electrode materials that facilitate the electroanalysis of HMIs due to their appealing characteristics such as low toxicity, ease of alloy formation with HMIs, and favorable electrochemical stripping behavior [23–25]. Compared to bismuth-based films, Bi nanoparticles (BiNPs) exhibit enhanced electrocatalytic activity attributed to their large specific surface areas along with abundant active sites [26]. Furthermore, Bi(III) cluster-based metal-organic compounds (MOCs) possess regular morphologies, diverse structures, tunable pore sizes, and rich active sites, which significantly enhance their applicability in this field [27–29]. The aforementioned structural characteristics promote the enrichment of HMIs to elevated levels, thereby realizing the electrochemical sensing of HMIs with high sensitivity [30,31]. Therefore, Bi(III) cluster-based MOCs are anticipated to represent a promising class of sensing materials for the detection of Cd^{2+} and Pb^{2+} .

In this work, a dumbbell-shaped tetranuclear Bi(III) cluster $[\text{Bi}_4(\text{TC4A})_2(\mu_4\text{-O})]\cdot 2\text{NO}_3\cdot 3\text{DMF}\cdot 3\text{MeOH}$ (**Bi₄-TC4A**) and a flower-shaped octanuclear Bi(III) cluster $[\text{Bi}_8(\text{TC4A})_5(\text{en})_2(\mu_3\text{-OH})_2(\mu_4\text{-OH})_2]\cdot 5\text{DMF}\cdot 2\text{MeOH}$ (**Bi₈-TC4A**) were synthesized by using the bowl-shaped *p*-tert-butylthiacalix[4]arene ($\text{H}_4\text{TC4A}$) ligands and Bi(III) cations. Considering their rich Bi(III) cations, bowl-shaped cavities, and the presence of Lewis soft base groups (compounds containing S, N, and O), these materials emerge as promising candidates for the detection of HMIs through the enrichment and adsorption of Cd^{2+} and Pb^{2+} . As a result, both Bi(III) clusters, **Bi₄-TC4A** and **Bi₈-TC4A**, were employed as electrochemical sensors for detecting Cd^{2+} and Pb^{2+} with wide concentration ranges. Notably, during simultaneous detection of these two HMIs, **Bi₈-TC4A** exhibited remarkably low limits of detection at $1.4\text{ }\mu\text{g L}^{-1}$ for Cd^{2+} and $1.7\text{ }\mu\text{g L}^{-1}$ for Pb^{2+} . These values are much lower than the guideline thresholds established by World Health Organization (WHO). Furthermore, these limit values position our findings at the forefront among reported electrochemical sensors for Cd^{2+} and Pb^{2+} detection.

2. Experimental section

2.1. Synthesis of **Bi₄-TC4A**

$\text{Bi}(\text{NO}_3)_3\cdot 5\text{H}_2\text{O}$ (0.1 g, 0.2 mmol) and $\text{H}_4\text{TC4A}$ (0.072 g, 0.1 mmol) were added to the mixed solution of DMF and MeOH (5:1, 6 mL). Then, ethylenediamine (en, 0.2 mL) was slowly added to the stirred solution. The resulting mixture was transferred into a Teflon-lined reactor (15 mL), heated at $130\text{ }^\circ\text{C}$ for 3 days, and gradually cooled to room temperature. Green cubic crystals were obtained by filter and dried in air ($\sim 38\%$ yield based on $\text{H}_4\text{TC4A}$). Elemental analysis (%) Calcd for $[\text{Bi}_4(\text{TC4A})_2(\mu_4\text{-O})]\cdot 2\text{NO}_3\cdot 3\text{DMF}\cdot 3\text{MeOH}$: C 45.18, H 4.86, N 2.15, Found: C 45.31, H 4.62, N 2.33. IR (KBr, cm^{-1}): 2955(w), 1647(m), 1423(s), 1384(m), 1288(s), 1246(s), 1085(w), 873(w), 826(m), 730(m), 636(w), 611(w), 525(m).

2.2. Synthesis of **Bi₈-TC4A**

$\text{Bi}(\text{NO}_3)_3\cdot 5\text{H}_2\text{O}$ (0.1 g, 0.2 mmol) and $\text{H}_4\text{TC4A}$ (0.08 g, 0.11 mmol) were added to the mixed solution of DMF and MeOH (1:1, 10 mL), and

then ethylenediamine (en, 0.2 mL) and water (0.2 mL) were dropwise added to the stirring solution. After that, the mixture was put into a Teflon-lined reactor (15 mL) and heated at $130\text{ }^\circ\text{C}$ for 3 days. Green rod-shaped crystals were achieved ($\sim 33\%$ yield based on $\text{H}_4\text{TC4A}$). Elemental analysis (%) Calcd for $[\text{Bi}_8(\text{TC4A})_5(\text{en})_2(\mu_3\text{-OH})_2(\mu_4\text{-OH})_2]\cdot 5\text{DMF}\cdot 2\text{MeOH}$: C 45.20, H 4.83, N 2.14, Found: C 46.26, H 4.30, N 2.43. IR (KBr, cm^{-1}): 2956(w), 2102(w), 1427(m), 1250(m), 1081(w), 878(w), 819(w), 734(w), 497(s).

2.3. Fabrication of Bi(III) cluster based electrochemical sensors

Initially, a bare glassy carbon electrode (GCE) was polished using alumina slurries of 0.3 and $0.05\text{ }\mu\text{m}$ particle sizes, respectively, followed by thorough washing with alcohol and deionized water. The construction of the Bi(III) cluster-based electrochemical sensor proceeded as follows: the bulk Bi(III) cluster compounds were subjected to thorough grinding. Subsequently, 3.0 mg of the ground powder was dispersed in 1 mL of 0.5 wt % Nafion, followed by 15 min of ultrasonic treatment to yield a homogeneous suspension. Then, a $5\text{ }\mu\text{L}$ of the resulting 3.0 mg mL^{-1} **Bi₈-TC4A** suspension was drop-cast onto the surface of the freshly polished GCE. Finally, the modified electrode was dried under an infrared lamp (Scheme 1).

3. Results and discussion

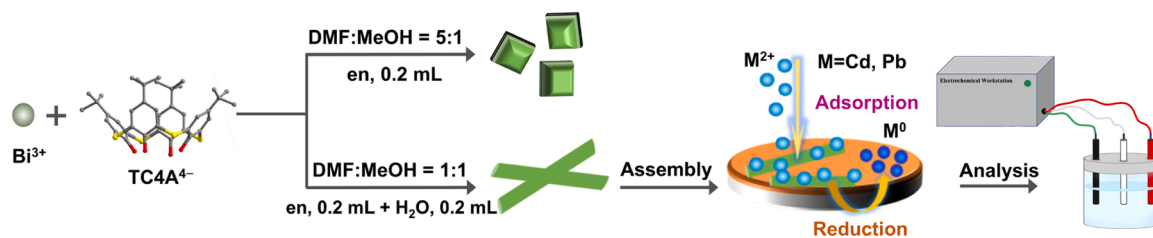
3.1. Crystal structures of **Bi₄-TC4A** and **Bi₈-TC4A**

Single-crystal X-ray diffraction analysis reveals that **Bi₄-TC4A** presents a dumbbell-shaped structure centered on a tetranuclear Bi(III) cluster (Fig. 1a). Specifically, its structural framework consists of a central $[\text{Bi}_4(\mu_4\text{-O})]^{10+}$ cluster capped by two TC4A^{4-} anions (Fig. 1c). In the **Bi₄-TC4A** architecture, the Bi1 atom exhibits a coordination environment involving four phenoxy O atoms, two S atoms from two TC4A^{4-} anions, and one bridging O atom derived from a coordinated water molecule (Fig. S1). The interatomic distances of Bi1–Bi1^{#1} and Bi1–Bi1^{#2} are measured as $3.215\text{ }\text{\AA}$ and $3.189\text{ }\text{\AA}$, respectively, which fall within the range of reported Bi–Bi bonding lengths ($2.947\text{--}3.219\text{ }\text{\AA}$) [32]. This observation allows the inference of metallic bonding interactions between adjacent Bi(III) cations.

Furthermore, by adjusting the solvent ratio, a new octanuclear cluster compound, **Bi₈-TC4A**, was successfully obtained. **Bi₈-TC4A** features a five-petal flower-shaped structure composed of a central $[\text{Bi}_8(\mu_3\text{-OH})_2(\mu_4\text{-OH})_2]^{20+}$ cluster and surrounded by five TC4A^{4-} anions (Fig. 1b). Among the eight Bi(III) cations in this structure, Bi1 and Bi2 share an identical seven-coordinated environment, being surrounded by four O atoms, one S atom, and two bridging OH^- groups; Bi7 and Bi8 adopt the same nine-coordinated configuration, involving four O atoms, two S atoms, two N atoms, and one bridging OH^- group; While the remaining Bi3, Bi4, Bi5 and Bi6 each exhibit an octahedral coordination sphere, coordinated with four oxygen atoms, two sulfur atoms, and two OH^- groups (Fig. S2). To conveniently distinguish five TC4A^{4-} anions in **Bi₈-TC4A**, the five anions are labeled as gray cup diagrams in Fig. 1d. Finally, the crystallographic data and the selected bond lengths as well as angles of **Bi₄-TC4A** and **Bi₈-TC4A** are shown in Tables S1–S3.

3.2. Characterization of **Bi₄-TC4A** and **Bi₈-TC4A**

We selected DMF and MeOH as the optimal solvent system and employed ethylenediamine as the chelating agent to innovatively synthesize bismuth cluster crystals through the solvothermal method. The size and morphology of the resulting crystals under different solvent ratios were systematically characterized using optical microscope and scanning electron microscopy (SEM). Fig. 2a and b present the SEM images of the assembled bismuth cluster crystals, with the corresponding insets showing their optical micrographs. The **Bi₄-TC4A** compound, which consists of the $[\text{Bi}_4(\mu_4\text{-O})]^{10+}$ cluster, forms diamond-shaped block



Scheme 1. Schematic representation of the synthesis process for $\text{Bi}_4\text{-TC4A}$ and $\text{Bi}_8\text{-TC4A}$, along with the electrochemical detection of Cd^{2+} and Pb^{2+} on $\text{Bi}_8\text{-TC4A}$ /GCE sensor ($\text{M} = \text{Cd}, \text{Pb}$).

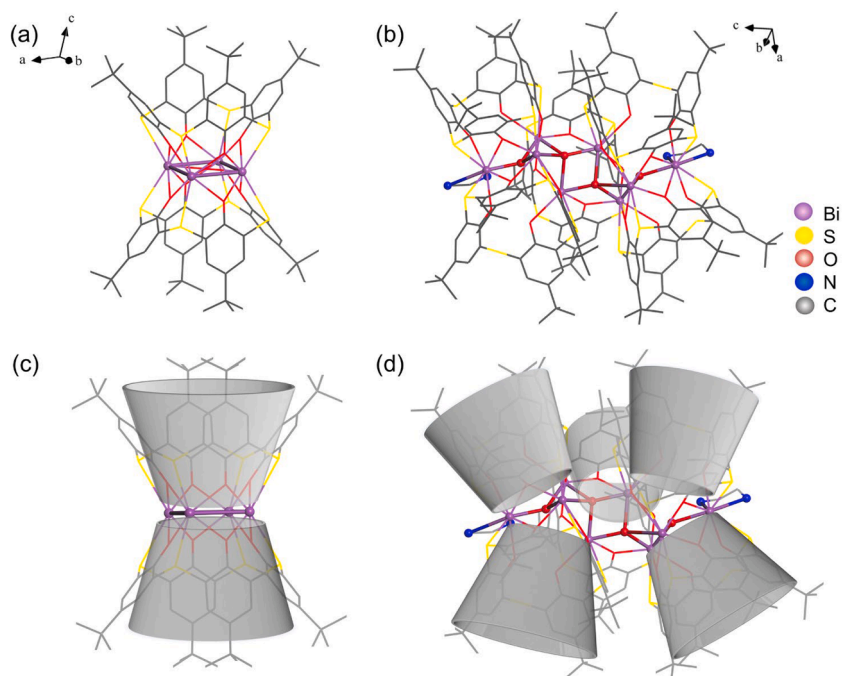


Fig. 1. Crystal structures of $\text{Bi}_4\text{-TC4A}$ (a) and $\text{Bi}_8\text{-TC4A}$ (b). Simplified coordination modes of $\text{Bi}_4\text{-TC4A}$ (c) and $\text{Bi}_8\text{-TC4A}$ (d).

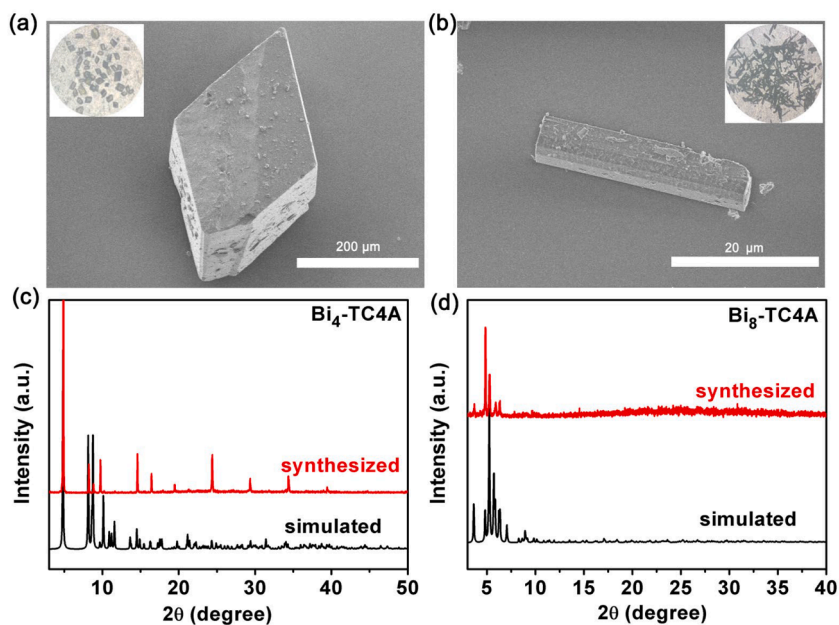


Fig. 2. SEM images of $\text{Bi}_4\text{-TC4A}$ (a) and $\text{Bi}_8\text{-TC4A}$ (b) (Insets: optical microscope photographs); The simulated and synthesized PXRD patterns for $\text{Bi}_4\text{-TC4A}$ (c) and $\text{Bi}_8\text{-TC4A}$ (d).

crystals measuring 200–400 μm in size. From the optical micrograph, these crystals are observed to be transparent with a pale green, displaying uniform thickness throughout (Fig. 2a). In contrast, **Bi₈-TC4A** compound, composed of the $[\text{Bi}_8(\mu_3\text{-OH})_2(\mu_4\text{-OH})_2]^{20+}$ cluster, crystallizes into dark green rod-like structures during slow solvent evaporation, with lengths ranging from 20 to 40 μm (Fig. 2b). Further, the phase purities of **Bi₄-TC4A** and **Bi₈-TC4A** were characterized by powder X-ray diffraction (PXRD) patterns (Fig. 2c and d). The PXRD patterns of the synthesized **Bi₄-TC4A** and **Bi₈-TC4A** matched well with the simulated ones, demonstrating the purity of the synthesized samples. To analyze the thermal stability of the two complexes, thermogravimetric (TG) curves of **Bi₄-TC4A** and **Bi₈-TC4A** were measured under nitrogen atmosphere from 25 to 800 $^{\circ}\text{C}$ (Fig. S3a and b). For **Bi₄-TC4A**, the weight loss before 299 $^{\circ}\text{C}$ can be ascribed to the release of three free DMF molecules and three free CH_3OH molecules. For **Bi₈-TC4A**, the weight loss before 312 $^{\circ}\text{C}$ corresponds to the release of five free DMF molecules and two free CH_3OH molecules.

Based on the results of the valence state calculations, **Bi₈-TC4A** is electrically neutral, whereas the framework of **Bi₄-TC4A** carries a positive charge. To achieve overall electrical neutrality in the compound, two positive charges from the framework were balanced by two NO_3^- anions, as evidenced by the absorption peak observed at approximately 1384 cm^{-1} in the FT-IR spectrum of **Bi₄-TC4A** (Fig. 3a) [33,34]. Furthermore, it is clear from the spectra that peaks at 1098 cm^{-1} and 556 cm^{-1} correspond to Bi-S and Bi-O bonds for both **Bi₄-TC4A** and **Bi₈-TC4A** (Fig. 3b) [35]. The broad peak around 500 cm^{-1} in **Bi₈-TC4A** arises from contributions of both Bi-N and Bi-O bonds [36].

Additionally, typical peaks located at 861 cm^{-1} are attributed to vibrations associated with Bi-O-Bi linkages [37]. Characteristic peaks found at 2960 cm^{-1} , 1656 cm^{-1} , 1450 cm^{-1} , and 1261 cm^{-1} can be assigned to C-H, C=C, $-\text{CH}_3$, and C-O groups present in calixarene respectively [38]. The presence of Bi(III) in **Bi₄-TC4A** and **Bi₈-TC4A** was confirmed through the analysis of the Bi 4f X-ray photoelectron spectroscopy (XPS) spectra (Figs. 3c and d and S4a and b). The spectrum for **Bi₄-TC4A** exhibited distinct peaks at binding energies of 164.1 eV and 158.8 eV, while that for **Bi₈-TC4A** displayed two unequal peaks at 163.8 eV and 158.5 eV, corresponding to the Bi 4f_{5/2} and 4f_{7/2} states, respectively. This observation indicates a separation of 5.3 eV, which is consistent with the oxidation state of Bi^{3+} [39]. Large specific surface areas and broad pore sizes are essential for effective adsorption applications [40,41]. The specific surface area and pore size distribution of the two compounds was examined using N_2 adsorption-desorption isotherms. As shown in Fig. 3e, the Brunauer-Emmett-Teller (BET) surface area for **Bi₈-TC4A** was calculated to be 141.13 $\text{m}^2 \text{g}^{-1}$, significantly higher than that of **Bi₄-TC4A**, which measured of 91.94 $\text{m}^2 \text{g}^{-1}$. Furthermore, the pore size distributions of **Bi₄-TC4A** and **Bi₈-TC4A** were analyzed from the desorption isotherm based on the Density Functional Theory (DFT) model, as illustrated in Fig. 3f. It is evident that both samples exhibit single-peak pore size distributions. The determined pore size was 3.55 nm for **Bi₄-TC4A** and 3.81 nm for **Bi₈-TC4A**.

3.3. Electrochemical performance of **Bi₄-TC4A**/GCE and **Bi₈-TC4A**/GCE

The electrochemical behavior of the electrolyte/electrode interface

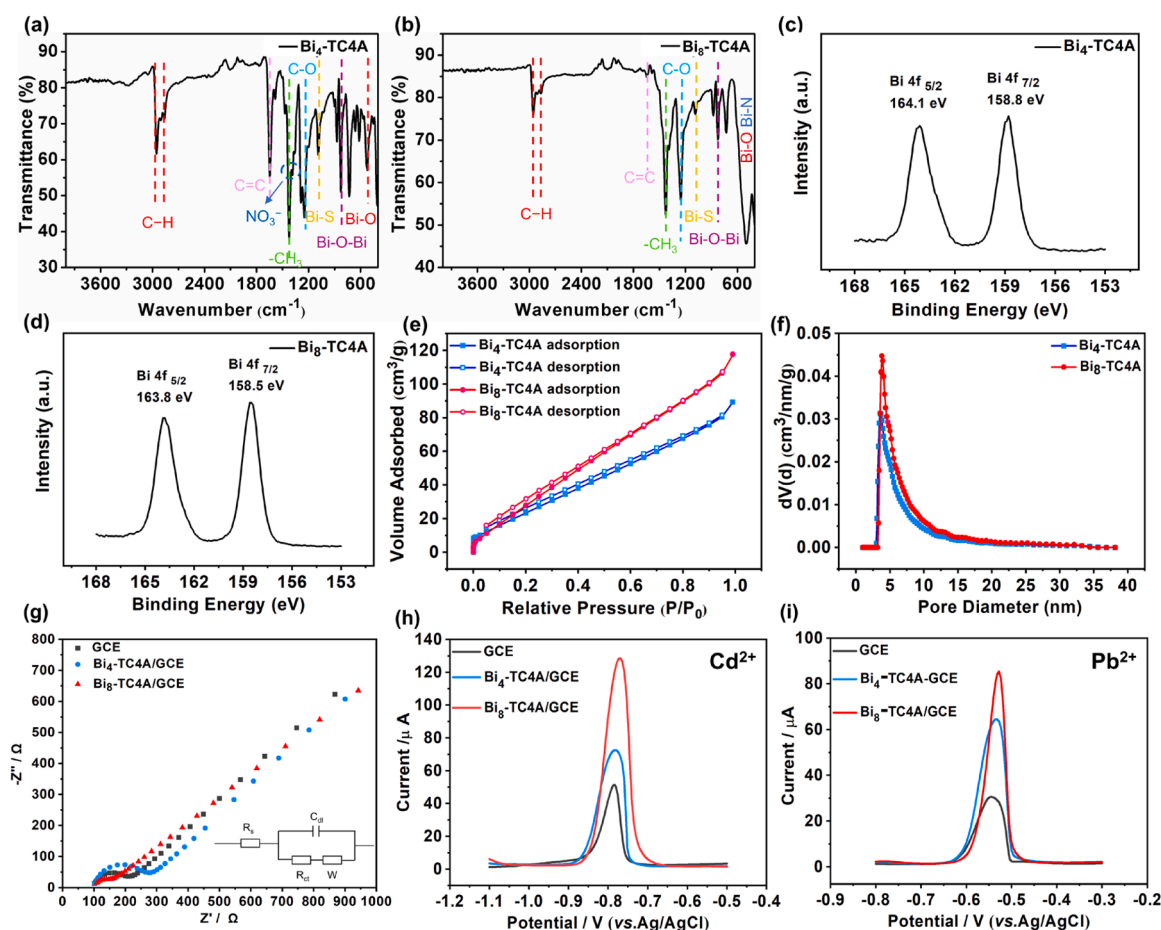


Fig. 3. FT-IR spectra of **Bi₄-TC4A** (a) and **Bi₈-TC4A** (b); XPS spectra of Bi 4f in **Bi₄-TC4A** (c) and **Bi₈-TC4A** (d); Plots of N_2 adsorption-desorption isotherms (e) and DFT pore size distributions (f) of **Bi₄-TC4A** and **Bi₈-TC4A**; Nyquist plots of bare GCE, **Bi₄-TC4A**/GCE and **Bi₈-TC4A**/GCE in the mixed solution of 10.0 mM $[\text{Fe}(\text{CN})_6]^{3-}/[\text{Fe}(\text{CN})_6]^{4-}$ (1:1) and 0.1 M KCl (g); SWASV responses of different electrodes in the presence of 1.0 $\text{mg}\cdot\text{L}^{-1}$ Cd^{2+} (h) and 1.0 $\text{mg}\cdot\text{L}^{-1}$ Pb^{2+} (i) in 0.1 M NaAC–HAC (pH = 5.5) buffer solution.

and the reaction kinetics on the **Bi₄-TC4A/GCE** and **Bi₈-TC4A/GCE** sensing platforms were systematically investigated using electrochemical impedance spectroscopy (EIS) in a mixed solution containing 10.0 mM $[\text{Fe}(\text{CN})_6]^{3-}/[\text{Fe}(\text{CN})_6]^{4-}$ (1:1) and 0.1 M KCl. As illustrated in Fig. 3g, Nyquist plots for different electrodes were recorded and fitted to an equivalent electrical circuit for data analysis (inset in Fig. 3g). In the Nyquist plots, the high-frequency intercept corresponds to the uncompensated solution resistance (R_s), which was found to be comparable across all tested samples. The charge-transfer resistance (R_{ct}), reflecting the electrochemical reaction kinetics in the middle- to low-frequency range, was used as a key parameter for comparison; lower R_{ct} values indicate faster electron transfer kinetics [42]. Notably, both **Bi₄-TC4A** and **Bi₈-TC4A** exhibited significantly smaller semicircular arcs compared to those typically conventional metal–organic framework materials (ZIF, UiO, PCN and MIL) [13,43–45]. The obtained R_{ct} values were 96.1 Ω for the bare GCE, 148.5 Ω for **Bi₄-TC4A/GCE**, and 55.3 Ω for **Bi₈-TC4A/GCE**. These results clearly demonstrate that **Bi₈-TC4A/GCE** not only reduces the semicircular arc radius but also enhances electron transfer efficiency during the electrochemical process, suggesting its potential as a highly effective electrochemical sensing platform for metal ions detection. Moreover, SWASV technology was employed to evaluate and compare the electrochemical detection performances of Cd^{2+} (1.0 $\text{mg}\cdot\text{L}^{-1}$) and Pb^{2+} (1.0 $\text{mg}\cdot\text{L}^{-1}$) on the bare GCE, **Bi₄-TC4A/GCE**, and **Bi₈-TC4A/GCE**, respectively. Compared with the bare GCE, significantly enhanced current signals for both Cd^{2+} and Pb^{2+} were observed on **Bi₄-TC4A/GCE** and **Bi₈-TC4A/GCE** (Fig. 3h and i). Notably, the peak currents obtained on **Bi₈-TC4A/GCE** were considerably higher than those on **Bi₄-TC4A/GCE**. This suggests that the abundant sulfur donor sites and structural cavities in **Bi₈-TC4A** facilitate the binding and accumulation of metal ions, thereby amplifying the electrochemical response. Consequently, **Bi₈-TC4A/GCE** exhibited more favorable sensing performance for detecting Cd^{2+} and Pb^{2+} .

3.4. Parameter optimization

During the electrochemical detection process, parameters such as electrolyte types, electrolyte pH value, deposition potential, and

deposition time play crucial roles in influencing sensing performance [46,47]. Typically, these parameters affect the conversion rate and efficiency from M^{2+} (where $\text{M} = \text{Cd}$ or Pb) to M^0 before being stripped back into solution [48]. The electrochemical detection method employed here involves the electro-deposition of M^{2+} followed by the anodic stripping of M^0 [49]. Consequently, we selected four buffer solutions including 0.2 M acetic acid buffer solution (ABS), 0.2 M phosphoric acid buffer solution (PBS), 0.2 M Britton–Robinson buffer solution (B–R), and 0.2 M NH_4Cl –HCl buffer solution for comparison the stripping peak currents of Cd^{2+} and Pb^{2+} at the **Bi₈-TC4A/GCE** (Fig. S5). Distinct discrepancies in the electrochemical responses were observed across the tested electrolyte systems. Specifically, in PBS and B–R buffer, the stripping peak current of Cd^{2+} was negligible, while that of Pb^{2+} was nearly undetectable. In the NH_4Cl –HCl buffer system, both Cd^{2+} and Pb^{2+} yielded measurable response currents, but the peak intensities were relatively low accompanied by irregular peak profiles. By contrast, well-resolved and high-intensity stripping peaks for both heavy metal ions were exclusively achieved in the ABS buffer solution. On this basis, ABS was identified as the optimal electrolyte for the subsequent detection of Cd^{2+} and Pb^{2+} . Subsequently, we investigated the effect of pH on peak current within a range of 4.0 to 6.0 (Fig. 4a and d). As pH values increased from 4.0 to 5.5, there was a significant enhancement in current signals due to strong adsorption and coordination of M^{2+} into **Bi₈-TC4A**. However, at a higher pH value of 6.0, signal attenuation may be attributed to the hydrolysis of Cd^{2+} and Pb^{2+} ions [13,50]. Therefore, optimal electrochemical detection for Cd^{2+} and Pb^{2+} occurred at a pH value of 5.5. The influence of deposition potentials on oxidative signals for Cd^{2+} and Pb^{2+} was examined by varying deposition potentials from -1.0 V to -1.4 V on **Bi₈-TC4A/GCE** (Fig. 4b and e). Stripping signals for both ions increased with more negative deposition potentials and the maximum peak currents were observed at -1.3 V, which was subsequently designated as the standard deposition potential for further tests. Similarly, we assessed how deposition time affected oxidation signals for Cd^{2+} and Pb^{2+} (Fig. 4c and f), maintaining a constant deposition potential at -1.3 V. The stripping currents for both Cd^{2+} and Pb^{2+} showed an increase with extended deposition times up to 300 s. Beyond this point, there was a notable decrease in current for Cd^{2+} , while that for

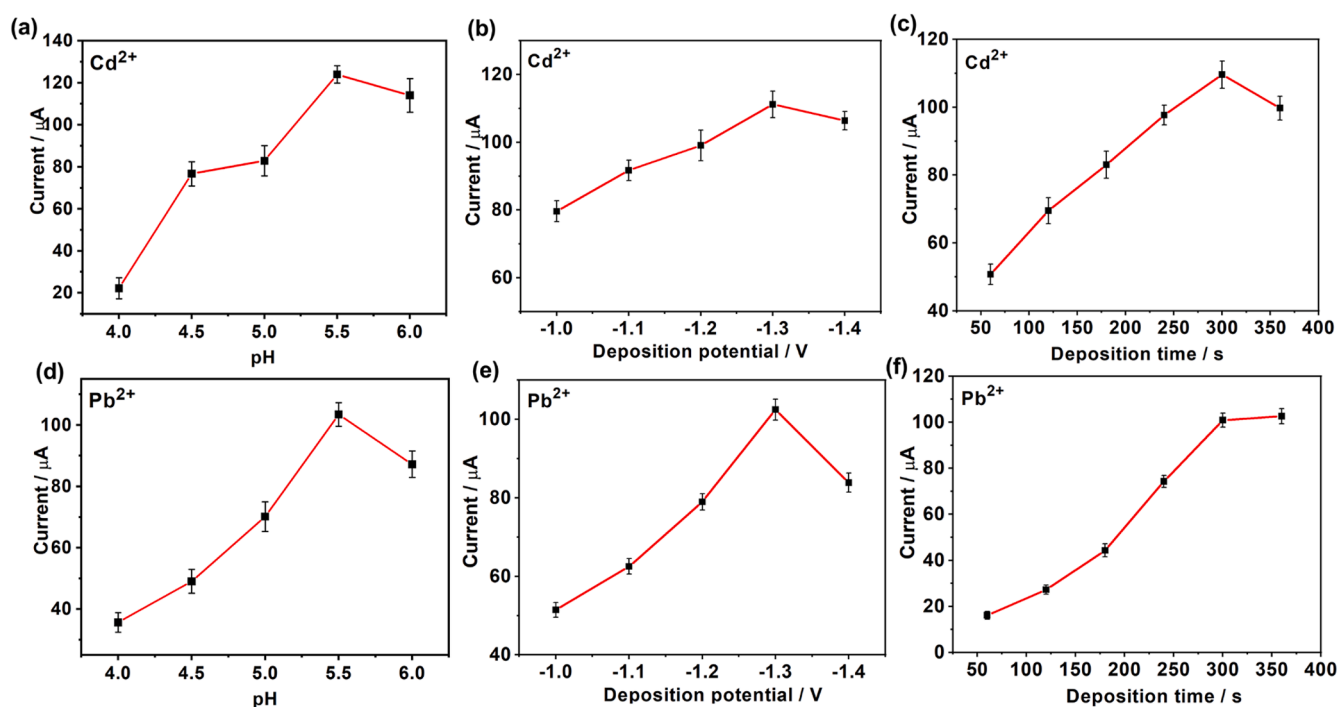


Fig. 4. Effect of the supporting electrolyte pH values (a, d), deposition potentials (b, e) and deposition time (c, f) on the current responses of 1.0 $\text{mg}\cdot\text{L}^{-1}$ Cd^{2+} and 1.0 $\text{mg}\cdot\text{L}^{-1}$ Pb^{2+} at **Bi₈-TC4A/GCE**.

Pb^{2+} remained stable. Consequently, a standard deposition time of 300 s would be utilized in subsequent electroanalytical studies.

3.5. Electrochemical determination of Cd^{2+} and Pb^{2+}

In this study, **Bi₈-TC4A/GCE** was used to monitor Cd^{2+} and Pb^{2+} under the optimal conditions with 0.1 M NaAc–HAc buffer solution (pH = 5.5). As illustrated in Fig. 5a and c, the anodic stripping peaks towards Cd^{2+} and Pb^{2+} appeared at around -0.80 V and -0.55 V, respectively. Obviously, the separation of the stripping potentials made it possible to detect Cd^{2+} and Pb^{2+} simultaneously in the mixed solutions. The oxidation peak currents of Cd^{2+} increased gradually with the raised Cd^{2+} concentrations. The resulting calibration data plots were found to be linearly proportional to the concentrations over the range of 0.01 – 1.1 $\text{mg}\cdot\text{L}^{-1}$. Particularly, the relationship of the Cd^{2+} concentrations and peak current values could be expressed as $I(\mu\text{A}) = 113.08 C(\text{mg}\cdot\text{L}^{-1}) - 0.68$, with a high-correlation coefficient (R^2) value of 0.9939 fitted from Fig. 5b. The limit of detection (LOD) was 1.9 $\mu\text{g}\cdot\text{L}^{-1}$, which is absolutely effective to identify Cd^{2+} within the maximum allowable value (0.003 $\text{mg}\cdot\text{L}^{-1}$) set by World Health Organization (WHO) [51]. Similarly, the calibration curve for Pb^{2+} showed a good linear

relationship in the concentration range of 0.01 – 1.4 $\text{mg}\cdot\text{L}^{-1}$ (Fig. 5d), corresponding to the regression equation of $I(\mu\text{A}) = 87.54 C(\text{mg}\cdot\text{L}^{-1}) - 2.37$ ($R^2 = 0.9924$). The LOD was calculated to be 2.8 $\mu\text{g}\cdot\text{L}^{-1}$, which is much lower than the guideline value (0.01 $\text{mg}\cdot\text{L}^{-1}$) specified by WHO [52]. In the present study, the LOD was calculated following the formula $\text{LOD} = 3\sigma/b$, where σ denotes the standard deviation of the peak currents derived from 11 replicate measurements in a blank NaAc–HAc buffer solution, and b corresponds to the slope of the calibration plot.

Finally, the electrochemical sensing performances of Cd^{2+} and Pb^{2+} were measured simultaneously under the optimized conditions on **Bi₈-TC4A/GCE**. Two well-defined peaks at around -0.80 V and -0.55 V were observed for Cd^{2+} and Pb^{2+} , respectively (Fig. 5e). As illustrated in Fig. 5f, the stripping currents were enhanced with the increased analyte concentrations. The good linear relationships between the stripping currents and the concentrations of Cd^{2+} and Pb^{2+} were observed in the detection range from 0.005 to 1.2 $\text{mg}\cdot\text{L}^{-1}$, and linear equations are $I(\mu\text{A}) = 108.40 C(\text{mg}\cdot\text{L}^{-1}) + 2.94$ ($R^2 = 0.9917$) for Cd^{2+} and $I(\mu\text{A}) = 83.42 C(\text{mg}\cdot\text{L}^{-1}) + 0.78$ ($R^2 = 0.9983$) for Pb^{2+} . Compared with other bi-based material or nanomaterial electrodes (Table S4), the designed **Bi₈-TC4A/GCE** sensor exhibited satisfactory detection results. Relative to the previously reported sensors, **Bi₈-TC4A/GCE** exhibited extremely

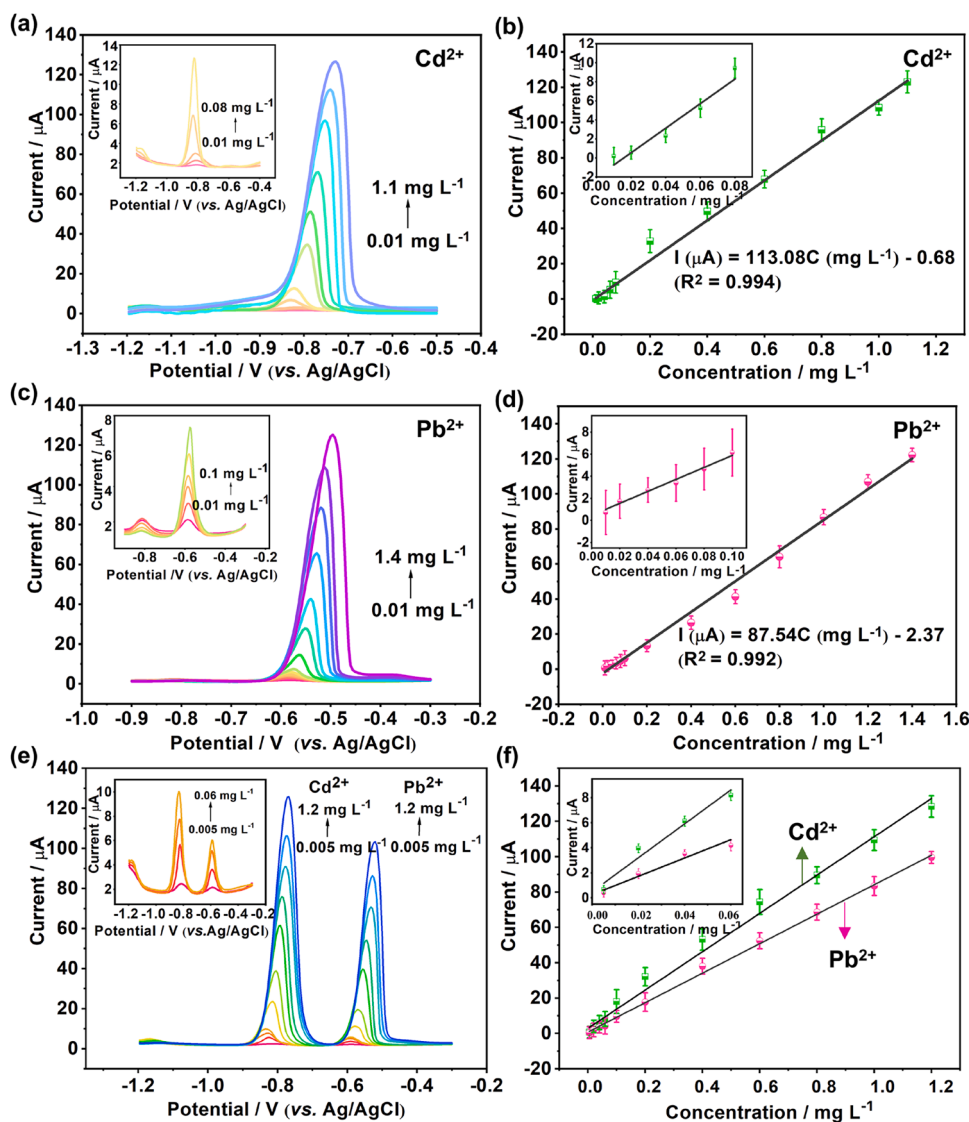


Fig. 5. Typical SWASV responses for Cd^{2+} and Pb^{2+} on **Bi₈-TC4A/GCE** (a, c) (Inset: the SWASV curves for low concentrations of Cd^{2+} or Pb^{2+}) and the corresponding linear calibration plots (b, d) (Inset: the magnifying linear calibration plots at low concentration of Cd^{2+} or Pb^{2+}); (e) SWASV responses for simultaneous detection of Cd^{2+} and Pb^{2+} over the concentration range from 0.005 to 1.2 $\text{mg}\cdot\text{L}^{-1}$; (f) The corresponding calibration plots toward Cd^{2+} and Pb^{2+} , respectively.

wide linear ranges for the detection of Cd^{2+} and Pb^{2+} , which spanned three orders of magnitude. Particularly, the LOD values for Cd^{2+} and Pb^{2+} were $1.4 \mu\text{g}\cdot\text{L}^{-1}$ and $1.7 \mu\text{g}\cdot\text{L}^{-1}$, respectively, which were at the upstream detection level in Cd^{2+} and Pb^{2+} sensors. Particularly, the LOD values are much lower than the standard drinking water ones from the WHO. More significantly, the constructed **Bi₈-TC4A/GCE** electrode is inexpensive and easy to fabricate than the known Bi film electrodes.

3.6. Mechanistic considerations

To get insights into the mechanism of first-class adsorption capacity and outstanding electrochemical performance to Cd^{2+} and Pb^{2+} on **Bi₈-TC4A/GCE**, we apply the DFT calculations to investigate the adsorption behaviors of Cd^{2+} and Pb^{2+} on **Bi₈-TC4A** surfaces focusing on both electronic structures and optimized chemical bond parameters. The analysis of the density of states (DOS) indicates that although the Bi orbital projected density of states (PDOS) of both **Bi₄-TC4A** and **Bi₈-TC4A** cross the Fermi level (E_F), there is a notable difference in their electronic density of states distribution at the E_F (Fig. 6a and b). **Bi₈-TC4A** shows a clear DOS peak at E_F , whereas the DOS for **Bi₄-TC4A** near E_F is approximately zero. This disparity in electronic structure accounts for the superior conductivity observed in **Bi₈-TC4A** compared to **Bi₄-TC4A** [53], which aligns well with results obtained from EIS tests. To further elucidate the electronic interactions between Cd^{2+} or Pb^{2+} ions and **Bi₈-TC4A** surfaces, we compared the partial PDOS before and after the adsorption of Cd^{2+} or Pb^{2+} at the most stable metal-top sites (Fig. 6c and d). For $\text{Pb}(\text{H}_2\text{O})_4^{2+}$, it is noteworthy that the dominant peak of Pb orbitals is situated significantly above the E_F . Following adsorption on **Bi₈-TC4A** surfaces, the corresponding s and p orbitals exhibit a pronounced downshift towards the E_F , which results from substantial electron transfer from **Bi₈-TC4A** [54]. In examining **Bi₈-TC4A** surfaces, we observe a favorable alignment between S-3p orbitals and Pb-6 s orbitals. This observation substantiates the formation of Pb-S bonds due to chemisorption. Furthermore, overlapping p-p orbital interactions between Pb and S sites are also detected, reinforcing efficient electron transfer mechanisms. The effective coupling between s-p and p-p orbitals ensures robust formation of Pb-S bonds. Similarly, it has been confirmed that S atoms exert a comparable adsorption effect on Cd^{2+} ions. In addition, DFT calculations also indicate that the aromatic ring structure can serve as binding sites for Cd^{2+} and Pb^{2+} ions, facilitating their adsorption through the well-known cation- π interaction. The optimized binding distances for the cation- π interactions are presented

in Fig. 6e-h, and the geometric parameters are detailed in Table 1. Considered that **Bi₈-TC4A** comprises five TC4A^{4-} anions and $[\text{Bi}_8(\mu_3\text{-OH})_2(\mu_4\text{-OH})_2]^{20+}$ cations, one of the TC4A^{4-} anions along with its directly connected atoms has been selected as the calculation model to enhance computational efficiency.

3.7. Reproducibility, storage stability and selectivity

To explore the reproducibility of **Bi₈-TC4A** modified electrode, five parallel experiments were conducted for the electrochemical detection of Cd^{2+} and Pb^{2+} . The SWASV responses for Cd^{2+} ($1.0 \text{ mg}\cdot\text{L}^{-1}$) and Pb^{2+} ($1.0 \text{ mg}\cdot\text{L}^{-1}$) on five **Bi₈-TC4A/GCE** electrodes are shown in Figs. 7a and S6a. It can be observed that the stripping currents on **Bi₈-TC4A/GCE** were nearly constant and the relative standard deviations (RSDs) were calculated to be 2.73 % and 4.63 % for Cd^{2+} and Pb^{2+} , respectively. These results demonstrated that **Bi₈-TC4A/GCE** featured good reproducibility for the electrochemical determination of Cd^{2+} and Pb^{2+} . To verify its storage stability, the **Bi₈-TC4A/GCE** sensor was employed for monitoring the concentrations of Cd^{2+} and Pb^{2+} every day continuously for a week (the electrode was stored at 4°C when it was not used). Compared with its initial values, the peak currents of Cd^{2+} and Pb^{2+} did not decrease significantly, remaining 94.67 % and 92.77 % of the initial current values, respectively (Figs. 7b and S6b). Furthermore, interference ions, such as K^+ , Na^+ , Mg^{2+} , Zn^{2+} , Ca^{2+} , Co^{2+} , Mn^{2+} , Cr^{3+} , Fe^{3+} and Al^{3+} , were used to assess the selectivity of the electrochemical sensor. The concentration of interfering ions was $10 \text{ mg}\cdot\text{L}^{-1}$, which was 10-fold higher than that of Cd^{2+} or Pb^{2+} . As illustrated in Fig. 7c, the addition of these interference ions did not lead to the obvious current value changes of Cd^{2+} and Pb^{2+} , demonstrating good selectivity of the sensing platform for the simultaneous detection of Cd^{2+} and Pb^{2+} .

3.8. Applications to lake water, soil and food samples

The applicability performance of the **Bi₈-TC4A/GCE** sensor was examined in real lake water, soil and food samples due to the presence of the unknown interfering substances during Cd^{2+} and Pb^{2+} detection. Before determination, lake water, soil leaching and orange juice solutions were filtered through a $0.22 \mu\text{m}$ membrane to remove the suspended solids, and then the actual samples were diluted with 0.2 M NaAc-HAc buffer solution (pH = 5.5) in a ratio of 1:9. The spinach samples were dipped with cotton swabs on their surfaces and then added to the 0.2 M NaAc-HAc buffer solution (pH = 5.5). SWASV

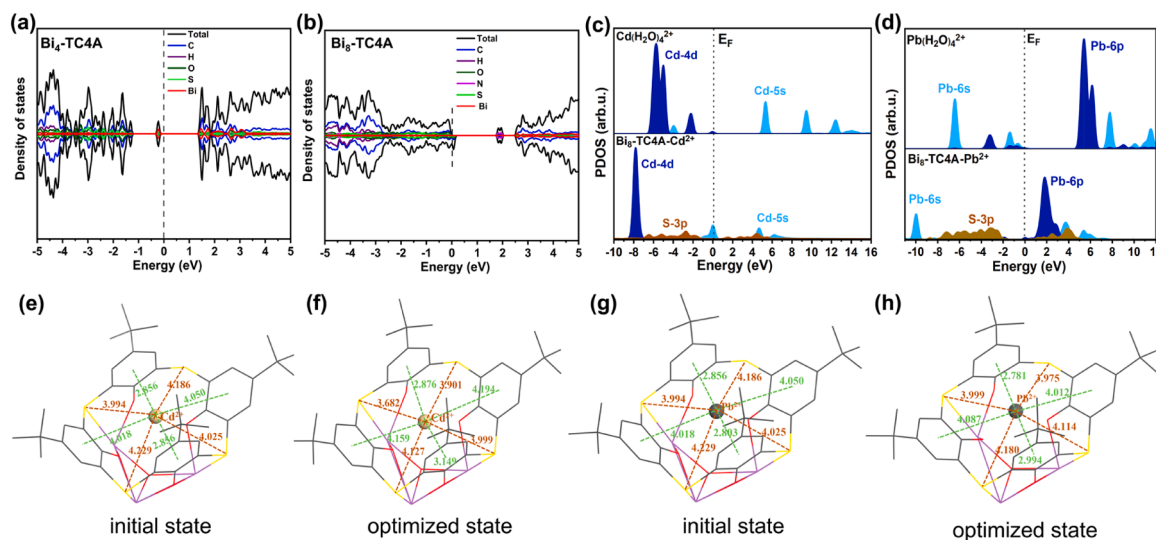


Fig. 6. Calculated DOS for (a) **Bi₄-TC4A** and (b) **Bi₈-TC4A**; The partial PDOS of comparisons for Cd^{2+} (c) or Pb^{2+} (d) adsorptions on **Bi₈-TC4A** (S-3p represent the 3p orbitals of S atoms of **Bi₈-TC4A**, $\text{Pb}(\text{H}_2\text{O})_4^{2+}$ represents hydrated lead ion, $\text{Cd}(\text{H}_2\text{O})_4^{2+}$ represents hydrated cadmium ion); Partial **Bi₈-TC4A** computational model structure (initial and optimized state for adsorption of Cd^{2+} and Pb^{2+}) and cation- π , Cd-S, and Pb-S binding distances (e-h).

Table 1
Optimized cation... π and M – S binding distances.

M ²⁺	Cation... π binding distances (Å)				M – S binding distances (Å)			
Cd ²⁺	2.876	4.194	3.149	4.159	3.901	3.999	4.127	3.682
Pb ²⁺	2.781	4.012	2.994	4.087	3.975	4.114	4.180	3.999

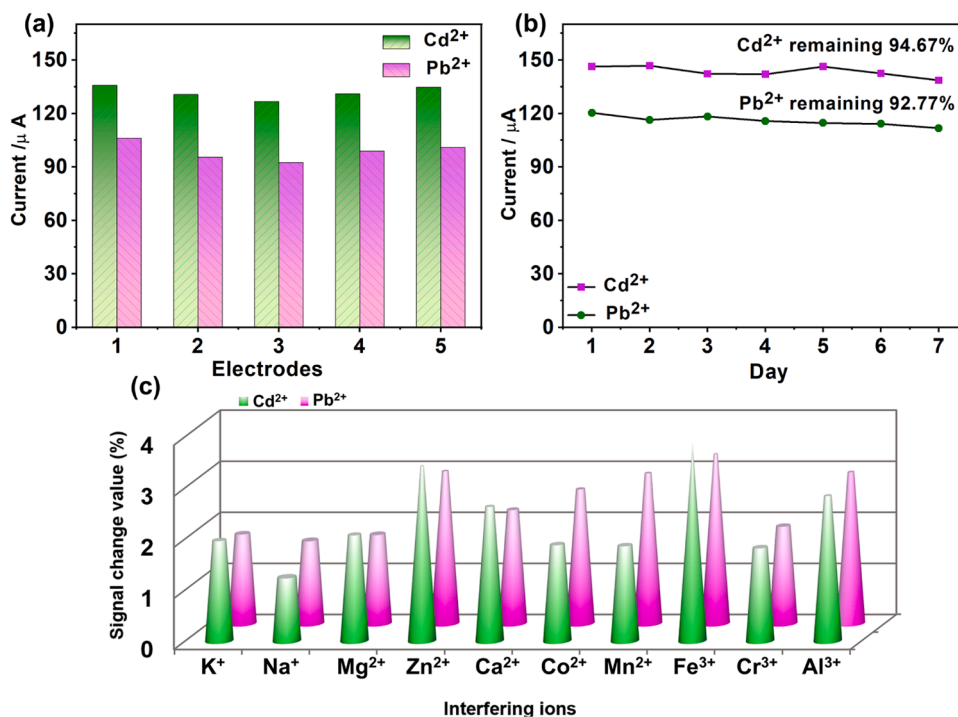


Fig. 7. (a) The repeatability of Bi₈-TC4A/GCE by the parallel determination of Cd²⁺ (mg·L⁻¹) and Pb²⁺ (mg·L⁻¹) for five times; (b) The stability of Bi₈-TC4A/GCE by the measurement of Cd²⁺ (mg·L⁻¹) and Pb²⁺ (mg·L⁻¹) every day for one week; (c) Percentage of current signal change values for Cd²⁺ (mg·L⁻¹) and Pb²⁺ (mg·L⁻¹) on the as-prepared Bi₈-TC4A/GCE in 0.1 M NaAc-HAc (pH = 5.5) containing 10-fold interfering ions.

determination results displayed that no current responses of Cd²⁺ and Pb²⁺ were observed in these samples, indicating that the concentrations of the target analytes were extremely low or no target analytes in the real samples. With the standard addition method, the practical application of the proposed sensor was evaluated by recovery determination of the spiked metal ion in water samples, and the results are listed in Table 2. It can be seen that the calculated average recovery ranged from 95.4 % to 104.0 %, demonstrating that the Bi₈-TC4A/GCE electrode featured a good potential application for detection of Cd²⁺ and Pb²⁺ in practical environment and food samples.

4. Conclusions

In summary, two novel Bi(III) clusters, Bi₄-TC4A and Bi₈-TC4A, were successfully synthesized by varying the ratio of the reaction solvents. Bi₄-TC4A exhibits a dumbbell-shaped tetranuclear structure, whereas Bi₈-TC4A features a flower-shaped octanuclear configuration. Compared to Bi₄-TC4A, Bi₈-TC4A demonstrates superior electrical conductivity and enhanced capacity for HMIs enrichment, rendering it an ideal candidate as an electrode modification material for the detection of Cd²⁺ and Pb²⁺. During electrochemical assessments of Cd²⁺ and Pb²⁺, the Bi₈-TC4A/GCE exhibited remarkable sensitivity, selectivity and stability. Notably, it holds significant potential for detecting Cd²⁺ and Pb²⁺ in real soil, lake water and food samples. The LODs for Cd²⁺ and Pb²⁺, determined using the Bi₈-TC4A/GCE system, are substantially lower than the guideline values established by WHO. The successful integration of calix[4]arene into both Bi(III)-based clusters provides an effective approach to enhance HMIs detection capabilities.

Table 2

Separate detection of Cd²⁺ and Pb²⁺ added in lake water and soil leaching solutions by SWASV.

Analytes	Samples	Original	Added (mg L ⁻¹)	Found ^a (mg L ⁻¹)	Recovery (%)	R.S.D (n = 3; %)
Cd ²⁺	Lake water	ND	0.050	0.048	96.0	3.11
			0.500	0.482	96.4	2.52
			0.050	0.051	102.0	2.88
Pb ²⁺	Soil	ND	0.500	0.503	100.6	1.15
			0.050	0.052	104.0	2.48
			0.500	0.511	102.2	3.47
Cd ²⁺	Lake water	ND	0.050	0.049	98.0	2.23
			0.500	0.520	104.0	1.98
			0.500	0.511	102.2	3.47
Pb ²⁺	Soil	ND	0.050	0.049	98.0	3.12
			0.500	0.515	103.0	2.98
			0.500	0.502	100.4	3.32
Cd ²⁺	spinach	ND	0.050	0.048	96.0	2.54
			0.500	0.477	95.4	2.88
			0.500	0.518	103.6	2.35
Pb ²⁺	orange juice	ND	0.050	0.049	98.0	3.12
			0.500	0.515	103.0	2.98
			0.500	0.502	100.4	3.32
Cd ²⁺	spinach	ND	0.050	0.048	96.0	2.54
			0.500	0.477	95.4	2.88
			0.500	0.518	103.6	2.35
Pb ²⁺	orange juice	ND	0.050	0.049	98.0	3.12
			0.500	0.515	103.0	2.98
			0.500	0.502	100.4	3.32

ND: no detection signal of analyte

^a the average of the three repeated tests.

CRediT authorship contribution statement

Chang Liu: Writing – original draft, Visualization, Data curation, Conceptualization. **Tianlong Shan:** Writing – review & editing, Validation. **Hongda Li:** Project administration, Funding acquisition. **Sitong Wu:** Writing – review & editing, Funding acquisition. **Meng Zhao:** Writing – review & editing. **Guannan Wang:** Methodology, Funding acquisition. **Rui Zhang:** Methodology, Funding acquisition, Conceptualization.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Supplementary materials

Supplementary material associated with this article can be found, in the online version, at [doi:10.1016/j.electacta.2025.148023](https://doi.org/10.1016/j.electacta.2025.148023).

Data availability

Data will be made available on request.

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